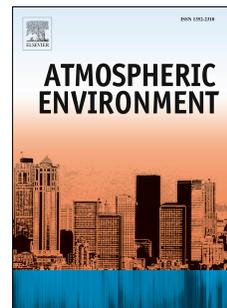


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The impact of biomass burning on total suspended particulate matter in the southeastern Tibetan Plateau

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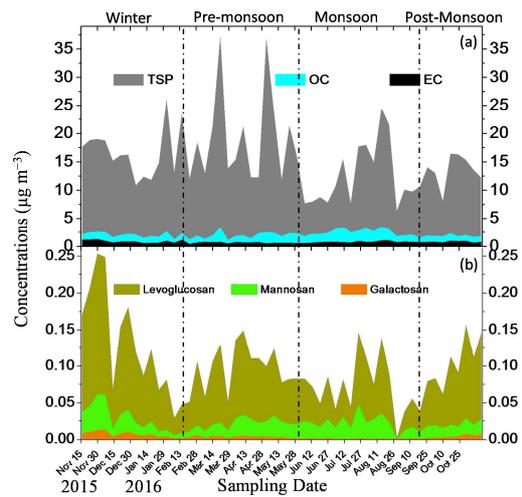
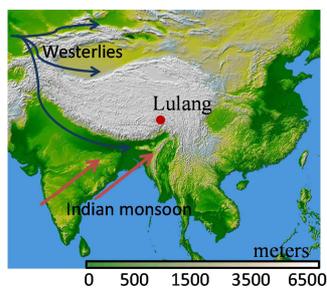
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14 **Abstract**

15 To investigate the impact of biomass burning (BB) emissions on total suspended
16 particulate (TSP) in the southeastern Tibetan Plateau (TP), BB tracers (including
17 levoglucosan, mannosan, and galactosan) were measured at Lulang (a remote site,
18 ~3300 m above sea level) during the period 2015–2016. The concentrations of
19 levoglucosan and mannosan showed large variability of more than 8-fold (range:
20 0.029–0.253 $\mu\text{g m}^{-3}$) and 6-fold (range: 0.01–0.061 $\mu\text{g m}^{-3}$), respectively. The highest
21 seasonal average concentrations of levoglucosan ($0.131 \pm 0.073 \mu\text{g m}^{-3}$) and
22 mannosan ($0.028 \pm 0.019 \mu\text{g m}^{-3}$) were observed during the winter season. The BB
23 contributions to the TSP organic carbon (OC) and TSP mass were estimated with
24 positive matrix factorization (PMF) by using BB tracers, inorganic ions, elements, OC,
25 and EC. The BB contributions to the TSP OC and TSP mass were quite substantial in
26 winter with 26% and 14%, respectively. The results evidenced a major contribution of
27 BB to the aerosol OC during winter season. Major potential source distributions of
28 BB were identified, which included the pollution bands along the Indo-Gangetic Plain,
29 the Yarlung Tsangpo River Valley and the interior of the TP. In addition to the
30 long-range transport of BB, the sources of anthropogenic emissions (residential BB
31 for cooking and heating) in the TP may be an important contributor. The results
32 facilitated understanding of the characteristics and effects of BB in the TP.

33

34 *Keywords: Levoglucosan; Biomass burning; Tibetan Plateau*

35

36

37 1. Introduction

38 The Tibetan Plateau (TP) plays important roles in global climate and environment
39 (Ramanathan et al., 2005, 2007; Yao et al., 2012). The TP glaciers with seasonal and
40 permanent snow cover affect the water supply of some Asian rivers. This has had a
41 significant impact on water resources for billions of local habitants in Asia (Ming et
42 al., 2013; Xu et al., 2009a). Previous studies found that carbonaceous aerosols can
43 reduce snow albedo, and are considered as important contributors to the accelerating
44 melting of Tibetan glaciers (Ming et al., 2013; Wang et al., 2015a; Xu et al., 2009a),
45 which is mainly from biomass and fossil fuel combustion emissions (Wang et al.,
46 2015b; Xu et al., 2009b; Xu et al., 2012; You et al., 2016).

47 Biomass burning (BB) aerosols are considered to be a major cause of the South
48 Asia Brown Clouds (Gustafsson et al., 2009; Engling and Gelencser, 2010). The BB
49 pollutants from the Indo-Gangetic Plain and South Asia can be transported to the
50 upper troposphere by thermal convection processes (Kahn et al., 2008; Vadrevu et al.,
51 2012), where they can be transported to the TP (Cong et al., 2015; Zhao et al., 2017).
52 Therefore, it is rather important and necessary to obtain insight into the impact of BB
53 aerosol in the above-mentioned vulnerable regions.

54 Levoglucosan is considered as a highly specific tracer for BB aerosols because it
55 can normally remain stable for several days under most atmospheric conditions during
56 the transport process (Mochida, 2003; Giannoni et al., 2012; Maenhaut et al., 2012;
57 Cong et al., 2015; Zangrando et al., 2016). Mannosan and galactosan can also be used
58 as BB tracers (Bernardoni et al., 2011; Maenhaut et al., 2012; Sang et al., 2013). The
59 use of levoglucosan as BB tracer has been preferred to the conventional water soluble
60 potassium, since potassium has other sources (Caseiro et al., 2009; Duan et al., 2004).
61 The study on the spatio-temporal characteristics of BB aerosols over the TP is still
62 limited due to its extremely high elevation and its harsh natural environment.

63 In this study, the variations of carbonaceous fractions and BB tracers (including
64 levoglucosan, mannosan, and galactosan) were investigated at a representative site
65 (Lulang, influenced by the Indian monsoon and westerly winds) in the southeastern
66 TP. The selection of the site could provide important information to better understand

67 the impact of air pollutants from the south of the Himalayas and the interior of the TP.
68 The BB contributions to total suspended particulate (TSP) mass and organic carbon
69 (OC) in the region were evaluated. Based on the BB tracer of levoglucosan, the major
70 potential source distributions of BB were obtained during the sampling period.

71

72 **2. Methods**

73 *2.1. Aerosol Sampling*

74 Figure 1 shows the sampling site location at Lulang (94.73°E, 29.76°N, 3326 m
75 a.s.l.) in the southeastern margin of the TP. In terms of meteorological conditions,
76 Lulang exhibits the typical characteristics of the Indo-Gangetic Plain climate with wet
77 monsoon season (June–September), dry winter season (December–February),
78 pre-monsoon season (March–May) and post-monsoon season (October–November).
79 The sampling was performed on the roof of a tower (10 m above the ground) in the
80 premises of the South-East Tibetan plateau Station for integrated observation and
81 research of alpine environment, Chinese Academy of Sciences (CAS).

82 TSP samples were collected on a weekly basis at 17 L/min with a custom-built
83 sampler on 47 mm Whatman quartz–fiber filters (QM/A) from November 2015 to
84 November 2016. Field blank filters were also collected periodically. Quartz–fiber
85 filters were pre-heated at 900 °C for 3 hours to remove the residual carbon. The
86 samples were stored immediately in a refrigerator at about –20 °C to prevent the
87 evaporation of volatile components after sampling.

88

89 Figure 1

90

91 *2.2. Measurement of OC and EC*

92 The quartz microfibre filters were analyzed for carbon fractions using a DRI Model
93 2001 Thermal/Optical Carbon Analyzer (Atmoslytic Inc., Calabasas, CA, USA).
94 Carbon fractions were obtained following the IMPROVE-A (Interagency Monitoring
95 of Protected Visual Environments) thermal/optical reflectance protocol (Chow et al.,
96 2007). The method reports data for four OC fractions in a helium atmosphere (OC1,

97 OC2, OC3, and OC4 at 140 °C, 280 °C, 480 °C, and 580 °C, respectively), a
98 pyrolyzed carbon fraction (OP), and three elemental carbon (EC) fractions in a 2%
99 oxygen/98% helium atmosphere (EC1, EC2, and EC3 at 580 °C, 740 °C, and 840 °C,
100 respectively). OC and EC were defined as OC1+OC2+OC3+OC4+OP and
101 EC1+EC2+EC3-OP, respectively. Field blank filters were also analyzed in order to
102 correct the results. The detailed procedures of quality assurance/quality control
103 (QA/QC) have been reported elsewhere (Cao et al., 2003; Chow et al., 2011).

104

105 *2.3. Measurement of levoglucosan, mannosan, and galactosan*

106 The TSP samples were analyzed for levoglucosan, mannosan, and galactosan by
107 using high-performance anion exchange chromatography with pulsed amperometric
108 detection (HPAEC-PAD). A detailed description of the analytical method has been
109 reported elsewhere (Engling et al., 2006). The HPAEC-PAD method has been
110 developed and validated for determination of atmospherically relevant sugar alcohols,
111 monosaccharides, and monosaccharide anhydrides (Zhang et al., 2013). The
112 instrumental limit of detection was determined as the peak height at least three times
113 the signal to noise ratio. The HPAEC-PAD method detection limits were 4 ppb for
114 levoglucosan and 2 ppb for mannosan and galactosan, respectively.

115

116 *2.4. Measurement of inorganic ions and elements*

117 Four anions (SO_4^{2-} , NO_3^- , Cl^- , and F^-) and five cations (Na^+ , NH_4^+ , K^+ , Mg^{2+} , and
118 Ca^{2+}) in aqueous extracts of the filters were analyzed using ion chromatography (IC).
119 The elemental concentrations were determined using energy-dispersive x-ray
120 fluorescence (ED-XRF) spectrometry, which provides a rapid method for the analysis
121 of trace and major elements in samples. A detailed description of the detection limits
122 and methodology for the analyses can be found elsewhere (Cao et al., 2012).

123

124 **3. Results and discussion**

125 *3.1. Variations of TSP, OC, and EC mass*

126 Figure 2a shows the variations of TSP, OC, and EC mass at the Lulang site.
127 Compared to OC and EC, the weekly TSP concentrations showed larger variations
128 during the sampling period. TSP concentrations varied over 5.8-fold from a low of
129 6.39 to $37.28 \mu\text{g m}^{-3}$, and OC varied over 3.2-fold from 1.11 to $3.52 \mu\text{g m}^{-3}$. The
130 pattern indicated that there were important seasonal contributors to TSP besides of
131 carbonaceous fraction.

132 Annual and seasonal average TSP, OC, and EC mass concentrations were reported
133 in Table 1 during the sampling period. TSP levels were higher in pre-monsoon (mean
134 value $19.1 \pm 8.3 \mu\text{g m}^{-3}$), with respect to winter ($16.7 \pm 4.4 \mu\text{g m}^{-3}$), monsoon ($12.5 \pm$
135 $5.5 \mu\text{g m}^{-3}$), and post-monsoon ($13.6 \pm 2.7 \mu\text{g m}^{-3}$), respectively. The highest OC
136 concentrations were observed in monsoon season ($2.52 \pm 0.61 \mu\text{g m}^{-3}$), followed by
137 winter ($2.1 \pm 0.44 \mu\text{g m}^{-3}$), pre-monsoon ($1.98 \pm 0.58 \mu\text{g m}^{-3}$), and post-monsoon
138 ($1.88 \pm 0.20 \mu\text{g m}^{-3}$). The annual concentrations of TSP, OC, and EC mass were 15.2
139 $\pm 6.4 \mu\text{g m}^{-3}$, $2.1 \pm 0.64 \mu\text{g m}^{-3}$, and $0.74 \pm 0.22 \mu\text{g m}^{-3}$, respectively. The variations
140 of the EC mass showed small difference during the sampling period (Figure 2a).

141

142 Table 1

143

144 3.2. The variations of levoglucosan, mannosan, and galactosan

145 The concentrations of levoglucosan and mannosan showed large variability during
146 the study period (Figure 2b). The weekly levoglucosan concentrations varied over
147 8.7-fold from 0.029 to $0.253 \mu\text{g m}^{-3}$, and mannosan varied over 6-fold from 0.010 to
148 $0.060 \mu\text{g m}^{-3}$. Levoglucosan and mannosan exhibited a similar seasonal cycle, with
149 the highest concentrations of $0.131 \pm 0.073 \mu\text{g m}^{-3}$ and $0.028 \pm 0.019 \mu\text{g m}^{-3}$ in
150 winter, respectively. The lowest values of levoglucosan ($0.072 \pm 0.038 \mu\text{g m}^{-3}$) and
151 mannosan ($0.021 \pm 0.005 \mu\text{g m}^{-3}$) were observed during the monsoon and
152 post-monsoon, respectively. Comparable seasonal levels of galactosan were found in
153 the range of $0.001 \pm 0.001 \mu\text{g m}^{-3}$ (monsoon) to $0.007 \pm 0.004 \mu\text{g m}^{-3}$ (winter) (Table
154 1). It is to be noted that the galactosan concentrations of some monsoon samples were
155 lower than the instrumental limit of detection. The annual average concentrations for

156 levoglucosan, mannosan, and galactosan in the present study were $0.097 \pm 0.051 \mu\text{g}$
157 m^{-3} , $0.023 \pm 0.012 \mu\text{g m}^{-3}$, and $0.004 \pm 0.003 \mu\text{g m}^{-3}$, respectively. The highest
158 levels were noted in winter for the BB tracers, thus indicating that BB is most intense
159 in the coldest season of the year.

160 The concentrations of the TSP, OC and EC mass, and levoglucosan at Lulang were
161 lower than those measured in the southern area of the TP. For example, higher levels
162 of TSP ($196 \pm 132 \mu\text{g m}^{-3}$), OC ($32.8 \pm 21.5 \mu\text{g m}^{-3}$), EC ($5.95 \pm 2.70 \mu\text{g m}^{-3}$), and
163 levoglucosan ($0.73 \pm 1.04 \mu\text{g m}^{-3}$) were obtained at Lumbini, Nepal (Wan et al.,
164 2017). The concentrations of levoglucosan at Lulang were much lower than at
165 BB-influenced sites in India with values reported for New Delhi ($1.98 \mu\text{g m}^{-3}$),
166 Raipur ($2.18 \mu\text{g m}^{-3}$) and Rajim ($2.26 \mu\text{g m}^{-3}$), respectively (Deshmukh et al., 2016;
167 Li et al., 2014; Nirmalkar et al., 2015). The latter sites in Nepal and India were close
168 to anthropogenic sources and were possibly influenced by local agricultural burning
169 or forest fires.

170
171 Figure 2

172 173 3.3. Source identification and apportionment by PMF

174 PMF (positive matrix factorization) is a receptor model, which is a mathematical
175 approach for quantifying the contributions of sources to samples based on the
176 chemical composition or source fingerprints (Paatero, 1997). PMF 5.0 was used for
177 the current analysis.

178 The source profiles for the base model 7-factor solution and the average percentage
179 apportionments of the various species to the factors are shown in Figure 3.
180 Considering the key species, the source association of the 7 factors seemed overall
181 quite logical. Factor 1 was dominated by OC, EC, Zn, Mn, and Fe, and is best
182 interpreted as combustion aerosol and traffic-related emissions. The second PMF
183 factor had high loadings for OC, Na^+ , K^+ , Cl^- , and Mg^{2+} , and this factor can most
184 readily be explained as a combination of secondary organic carbon and sea salt.

185 Factor 1 and factor 2 accounted for 14.9% and 14.4% of TSP mass, respectively. The
186 levels of NO_3^- , Cl^- , Ca^{2+} and Mg^{2+} in the profiles of the third factor were very high,
187 and this can be attributed to the formation of secondary nitrate. Factor 4 was
188 characterized by high loadings of elements such as Ba, Fe, Ca, Ti, and Mn, and this
189 factor can most readily be explained by road dust and fugitive dust. Factor 3 and
190 factor 4 accounted for 13.6% and 17.7% of TSP mass, respectively. Factor 5 was
191 determined to be representative of secondary sulfate since SO_4^{2-} and NH_4^+ were
192 highly loaded on this factor. Factor 6 was enriched in crustal elements (e.g., Ca, Fe, Ti,
193 K, Mg^{2+} , Ca^{2+} , and Cl^-), and this factor is most likely associated with mineral aerosol.
194 This factor was the largest contributor to TSP mass, accounting for 25.9% during the
195 sampling periods. The results suggest that mineral aerosol is the most important
196 source for the TSP mass at Lulang. Factor 7 appeared to represent BB. Loadings on
197 this profile were high for levoglucosan, galactosan, OC, and EC. This source
198 accounted for 7.9% of TSP mass for the sampling period.

199

200

Figure 3

201

202 3.4. Contributions from BB to the TSP OC and TSP mass

203 Figure 4 shows the time series of the percentages TSP OC and TSP mass from BB
204 at Lulang. The patterns in both figures were fairly similar to that observed in Figure 2
205 for the time series of levoglucosan. The seasonal ranges of 3–43%, 6–30%, 5–24%,
206 and 8–34% for % TSP OC from BB were obtained in winter, pre-monsoon, monsoon,
207 and post-monsoon, respectively, whereas those for the % TSP mass from BB were in
208 the ranges of 3–22%, 3–16%, 4–23%, and 4–21% for winter, pre-monsoon, monsoon,
209 and post-monsoon, respectively. The percentage contributions from BB to the TSP
210 OC and TSP mass showed a clear seasonal character. The seasonal averages for the %
211 TSP OC from BB were 26%, 17%, 12%, and 17% for winter, pre-monsoon, monsoon,
212 and post-monsoon, respectively, while those for the % TSP mass were 14%, 8%, 10%,
213 and 10% for winter, pre-monsoon, monsoon, and post-monsoon, respectively. The

214 average BB contributions were quite substantial in winter for the TSP OC and TSP
215 mass. The lowest percentages were obtained for the % TSP mass in pre-monsoon and
216 the % TSP OC mass in monsoon, respectively. The annual contributions from BB to
217 the TSP OC and TSP mass were 18% and 11% during the sampling period,
218 respectively. The sharp variations clearly illustrated the impact from BB (including
219 local emissions and long-range atmospheric transport) on the TSP OC and TSP mass,
220 especially in winter.

221 The contributions of BB to the TSP mass in the present study were comparable to
222 those of 14% for Swiss sites (Gianini et al., 2013), while the values were much lower
223 than those of 37% (annual average) and 50% (during winter) of the PM₁₀ mass for an
224 urban background site in northern Tuscany (Nava et al., 2015). The winter
225 contributions of BB to the TSP OC mass in the present study were comparable to
226 those of PM₁₀ aerosol in Flanders, Belgium (Maenhaut et al., 2016), while the levels
227 were higher than those of 6.5-11% in Hong Kong (Sang et al., 2011), and 16-28% in
228 Hainan (Zhang et al., 2012), respectively.

229

230 Figure 4

231

232 From scatter plots of the PMF-derived BB TSP OC mass and TSP mass versus
233 levoglucosan, we can arrive at conversion factors of 3.6 (R= 0.93) and 14.4 (R= 0.93),
234 respectively (Figure 5). The conversion factors were comparable to those in Flanders,
235 Belgium (Maenhaut et al., 2012). The values estimated in the present study were
236 much lower than the PMF-derived conversion factors of 9.7 and 22.6 for OC and
237 PM₁₀, respectively (Maenhaut et al., 2016). Larger factors for particulates were
238 reported by some previous studies in Europe, with values reported for The
239 Netherlands, Helsinki (24 ± 9) and Switzerland (ranging from 15.3 to 29),
240 respectively (Kos and Weijers, 2009; Saarnio et al., 2012; Gianini et al., 2013). The
241 differences can be partly attributed to the variations in the type of biofuel and the
242 burning conditions. It is recommended to use the conversion factors for deriving the
243 contribution from BB when making use of levoglucosan as single marker at the

244 remote site.

245

246

Figure 5

247

248 3.5. The transport and potential source regions of biomass burning

249 To better understand the transport mechanism and potential source areas of BB,
250 seasonal distributions of fire hot spots were acquired from MODIS observations
251 (<https://firms.modaps.eosdis.nasa.gov/>). In addition, to assess the impact of regional
252 transported air pollutants in the Lulang region, three-day backward air-mass
253 trajectories corresponding to each sampling date were calculated using the HYSPLIT
254 model (Draxler and Rolph, 2012). The potential source contribution function (PSCF)
255 has been used to identify the potential locations of regional pollution sources (Polissar
256 et al., 1999; Hsu et al., 2003). In this study, the PSCF values were calculated by using
257 levoglucosan and three-day back trajectories arriving at Lulang at 1000 LST (local
258 standard time) during the sampling period.

259 Figure 6 shows that the pre-monsoon period is the major fire season in the lowland
260 of the southern Himalayas. The fire spots were seen to be more concentrated in areas
261 in India and Nepal compared with the interior of the TP. A previous study also
262 reported that there were intense agricultural burning activities and forest fires along
263 the southern Himalayan foothills and the Northern Indo-Gangetic Plain in
264 pre-monsoon. The pollutants can enter the TP by the valleys in the eastern Himalayan
265 section (Cong et al., 2015).

266

267

Figure 6.

268

269 Figure 7 shows the trajectories that were calculated from the individual three-day
270 backward trajectories in four seasons. For discussion purposes, we arbitrarily defined
271 a trajectory as “polluted” (brown color) when the levoglucosan concentrations exceed
272 the seasonal geometric mean concentration plus one geometric standard deviation of
273 the mean. Of all backward trajectories included in the analysis for winter, ~21% of

274 those were considered polluted. The average levoglucosan concentration for these
275 polluted trajectories was $0.24 \mu\text{g m}^{-3}$. The air masses in the pre-monsoon cluster
276 originated from northern India and over central Tibet. The percent contribution of
277 ~20% for the trajectories in the pre-monsoon cluster were classified as polluted and
278 had a mean value of $0.14 \mu\text{g m}^{-3}$. The air masses grouped into the monsoon cluster
279 were from the northeastern India with a polluted trajectories contribution of 19%, and
280 the average levoglucosan mass concentration of the polluted trajectories was $0.13 \mu\text{g}$
281 m^{-3} . The air masses in post-monsoon originated from the Indo-Gangetic Plain and
282 then moved across northeastern India and to the southeast of Tibet before arriving at
283 Lulang. The average levoglucosan concentration for these polluted trajectories was
284 $0.15 \mu\text{g m}^{-3}$. The percentage of the polluted trajectories assigned to this cluster was
285 ~25%.

286 In the winter and pre-monsoon seasons, strong westerlies pass through Nepal,
287 northwest India and Pakistan as well as the interior area of Tibet. In the monsoon and
288 post-monsoon season, air masses are derived from Bangladesh and northeast India
289 and bring moisture that originates from the Bay of Bengal. The highest levoglucosan
290 level for the “polluted trajectory” was found in winter season. In addition to natural
291 fire emissions, anthropogenic emissions (residential BB for cooking and heating) may
292 be important BB sources in the TP.

293

294

Figure 7

295

296 Figure 8 shows the PSCF plots for the major BB source areas, in which PSCF
297 values were displayed with regard to color scale. The important potential source
298 regions (PS1, PS2, PS3, and PS4), were likely located in the high PSCF value regions.
299 PS1 covered the Yarlung Tsangpo River Valley and local areas. Air pollutants can be
300 transported through the Yarlung Tsangpo River Valley from the Indo-Gangetic Plain
301 and Bangladesh to the site. PS2 covered the interior of the TP from the western area
302 (including some villages and cities). Considering the scarce distribution of fire hot
303 spots in the area, the results for PS1 and PS2 suggest the influence of residential BB

304 from cooking and heating. PS3 and PS4 included the obvious pollution bands along
305 the Indo-Gangetic Plain, which was consistent with the fire hot spots area in
306 pre-monsoon and winter. The pollution bands covered the northern area of India,
307 western area of Nepal and the area of Katmandu. The prevailing wind direction
308 around the southeastern margin of the TP was southerly, and the pollutants in the
309 pollution bands could be transported to Lulang along the valley of the Yarlung
310 Tsangpo River as a “leaking wall” (Cao et al., 2011). The air mass transport from the
311 Indo-Gangetic Plain region to the TP has also been proposed to follow specific
312 pathways (Cong et al., 2015; Zhao et al., 2017).

313

314 Figure 8

315

316 **Conclusions**

317 In the present study, the variations of carbonaceous fractions and BB tracers
318 (including levoglucosan, mannosan, and galactosan) were investigated for TSP at
319 Lulang in the southeastern margin of the TP. The results provided a better
320 understanding on the magnitude of levoglucosan, mannosan, and galactosan. The
321 seasonal variation patterns of BB tracers are different from TSP and OC. The highest
322 seasonal concentrations for TSP, OC, and BB tracers (levoglucosan and mannosan)
323 were observed during pre-monsoon, monsoon, and winter seasons, respectively. The
324 annual contributions from BB to the TSP OC and TSP mass were 18% and 11%, and
325 were quite substantial in winter with 26% and 14%, respectively. The results indicated
326 that BB is an important contributor, especially in winter. The seasonal BB pollutant
327 transport and major potential source distributions of BB were identified, which
328 included the pollution bands along the Indo-Gangetic Plain, the Yarlung Tsangpo
329 River Valley and the interior of the TP. In addition to long-range transport of BB
330 aerosol, anthropogenic BB aerosol from cooking and heating may be an important
331 contributor in the TP. The results can facilitate the knowledge for the characteristics
332 and effects of BB in the TP and guide policy actions aimed to effectively mitigate
333 emissions.

334

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342

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Table 1. Average and standard deviation values for TSP, OC, EC, and biomass burning tracers (levoglucosan, mannosan, and galactosan) concentrations ($\mu\text{g m}^{-3}$) across winter, pre-monsoon, monsoon, and post-monsoon during November 2015 to November 2016.

	TSP	OC	EC	levoglucosan	mannosan	galactosan
Annual	15.2±6.4	2.1±0.64	0.74±0.22	0.097±0.051	0.023±0.012	0.004±0.003
Winter	16.7±4.4	2.1±0.44	0.87±0.24	0.131±0.073	0.028±0.019	0.007±0.004
Pre-monsoon	19.1±8.3	1.98±0.58	0.63±0.11	0.098±0.031	0.022±0.007	0.003±0.001
Monsoon	12.5±5.5	2.52±0.61	0.78±0.14	0.072±0.038	0.023±0.010	0.001±0.001
Post-monsoon	13.6±2.7	1.88±0.20	0.81±0.13	0.105±0.033	0.021±0.005	0.004±0.002

1 Figure Captions:

2 Figure 1. Geographic location of the sampling site, Lulang.

3 Figure 2. Temporal variability of TSP mass, OC, and EC (a), and variations of
4 levoglucosan, mannosan, and galactosan for TSP (b).

5 Figure 3. Source profiles for TSP by PMF receptor model.

6 Figure 4. Time series of the % contribution from BB to the TSP OC and TSP mass at
7 Lulang.

8 Figure 5. Scatter plots of PMF-derived BB TSP mass and BB TSP OC vs.
9 levoglucosan and regression lines, forced through the origin.

10 Figure 6. Seasonal distributions of fire hot spots from the Terra and Aqua satellites
11 during the sampling period.

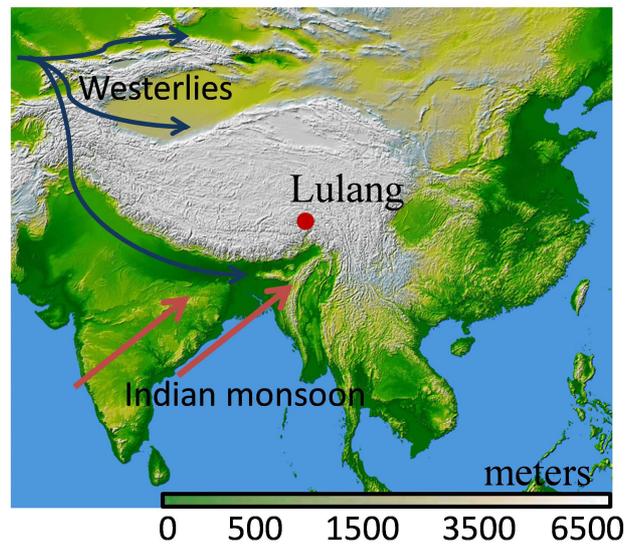
12 Figure 7. The seasonal three-day air mass backward trajectories with levoglucosan
13 concentrations using HYSPLIT model.

14 Figure 8. The four defined potential BB source distribution of PSCF using
15 levoglucosan concentrations.

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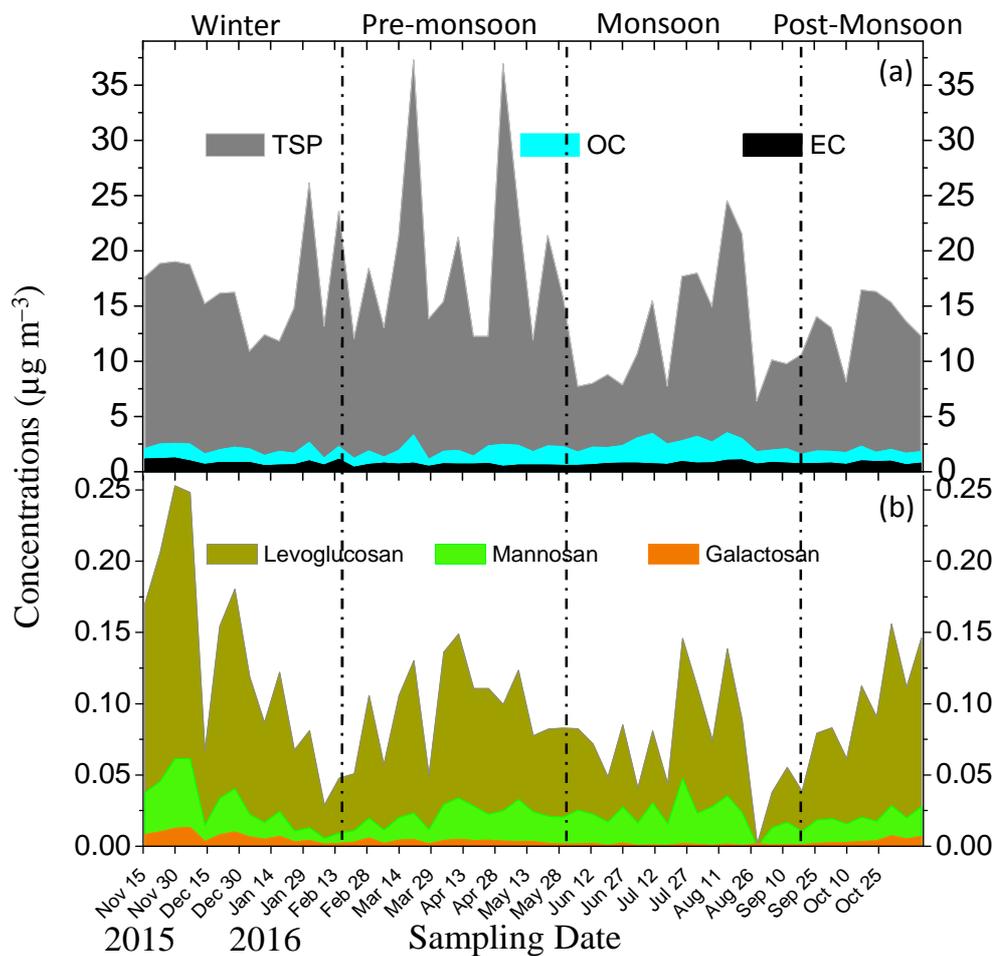


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20 Figure 1

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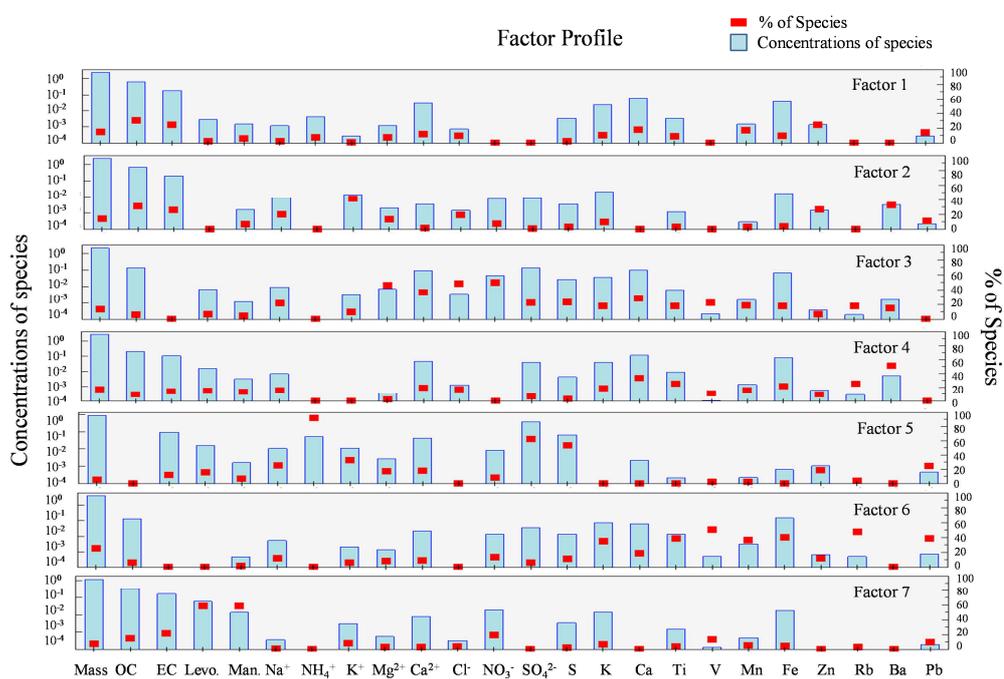
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24 Figure 2.

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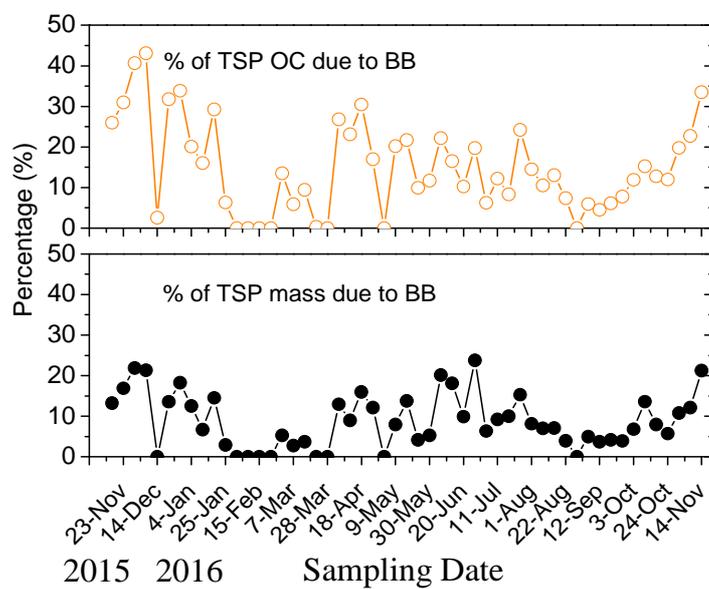
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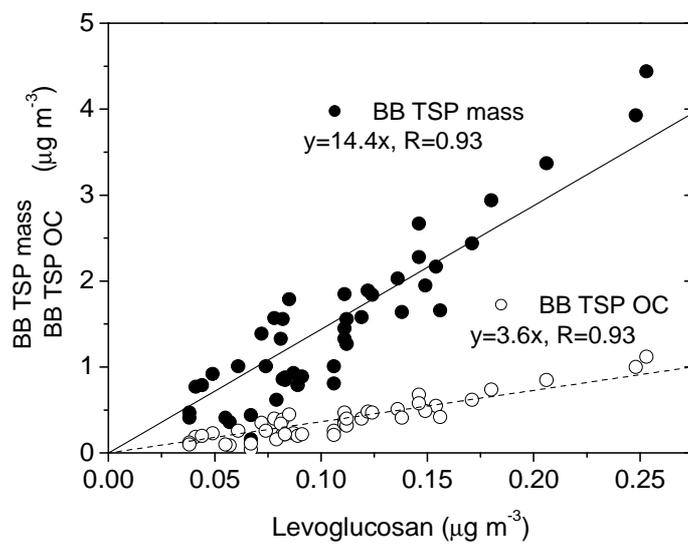
36 Figure 4.

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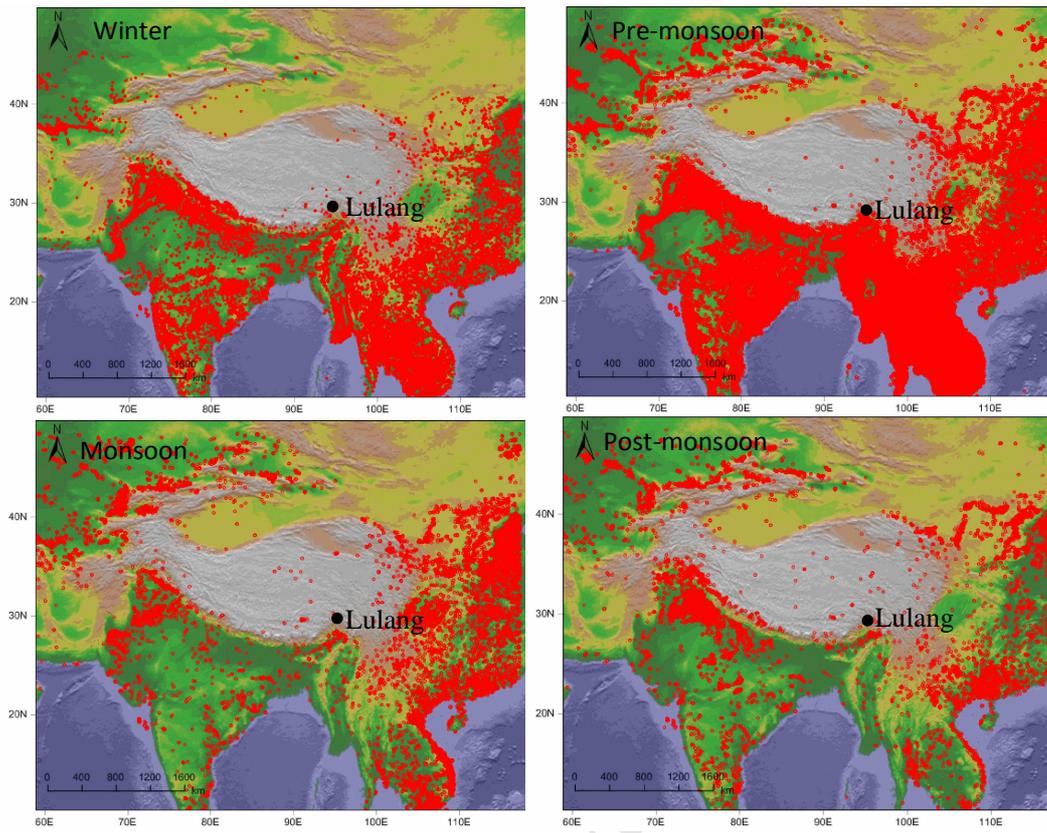
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42 Figure 5

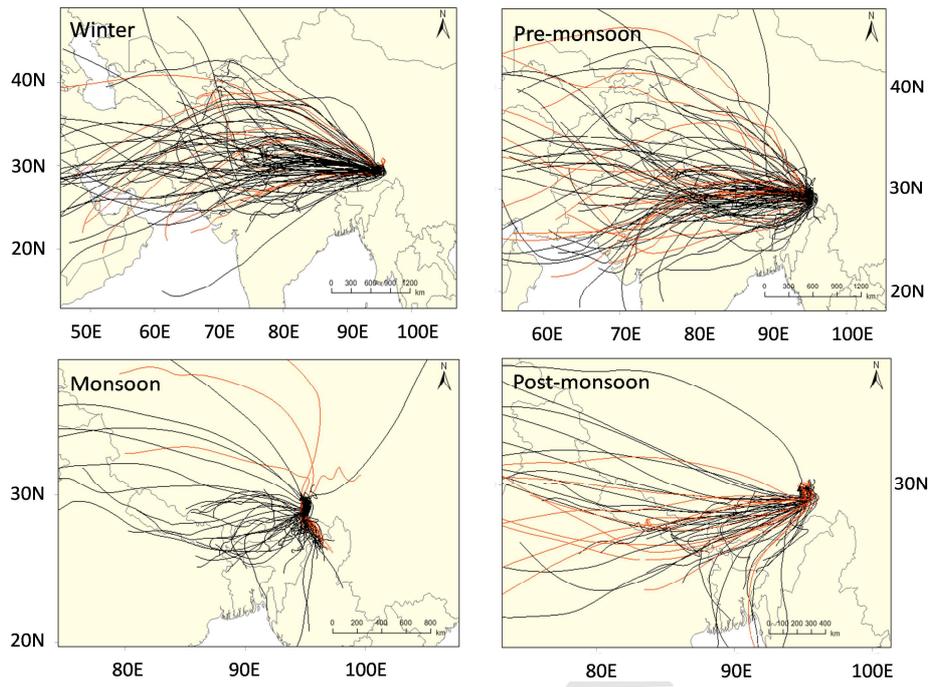
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45 Figure 6

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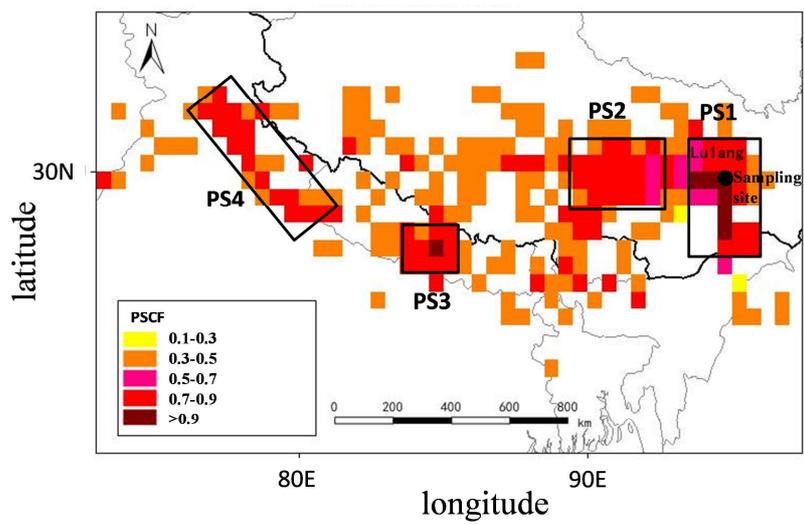


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Highlights:

1. The variations of carbonaceous fractions and BB tracers were investigated at Lulang in the southeastern TP.
2. The contributions from BB to the TSP and OC mass indicated that BB is an important contributor in winter.
3. The BB pollutant transport and major potential source distributions were identified.